DEVELOPMENT OF A SYSTEMWIDE PREDATOR CONTROL PROGRAM: STEPWISE IMPLEMENTATION OF A PREDATION INDEX, PREDATOR CONTROL FISHERIES, AND EVALUATION PLAN IN THE COLUMBIA RIVER BASIN

SECTION II: EVALUATION

1996 ANNUAL REPORT

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In Cooperation With

Washington Department of Fish and Wildlife
Pacific States Marine Fisheries Commission
Columbia River Inter-tribal Fish Commission
Confederated Tribes of the Umatilla Indian Reservation
Nez Perce Tribe of Idaho
Yakama Indian Nation
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Report E

Development of a Systemwide Predator Control Program: Indexing and Fisheries Evaluation

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ABSTRACT

Predator control fisheries aimed at reducing predation on juvenile salmonids by northern squawfish *Ptychocheilus oregonensis* were implemented for the seventh consecutive year in the mainstem Columbia and Snake rivers. In this report, we (1) evaluate northern squawfish exploitation and size composition, and compare catch rate of incidentally-harvested fishes among the three major management fisheries in 1996, (2) estimate reductions in predation on juvenile salmonids since implementation of the fisheries, and (3) evaluate changes from 1990-96 in relative abundance, consumption, size and age structure, growth, and fecundity of northern squawfish.

Systemwide exploitation of northern squawfish ≥ 250 mm fork length was 12.1% for sport-reward, 0.3% for dam-angling, and 0.5% for site-specific gill-net fisheries. Total exploitation was lowest in Lower Monumental Reservoir (0.0%) and highest in McNary Reservoir (18.2%). Mean fork length of harvested northern squawfish was 355 mm in the sport-reward, 391 mm in the dam-angling, and 408 mm in the gill-net fisheries. The dam-angling fishery had the lowest percentage (3.6%) of incidental catch relative to the total number of fish caught. Incidental catch was 27.0% in the sport-reward fishery and 54.6% in the gill-net fishery.

If exploitation rates remain similar to mean 1991-96 levels, we estimate that potential predation by northern squawfish on juvenile salmonids in 1997 will be approximately 62% of

predation levels prior to the implementation of removal fisheries. Further reductions in predation may be small, unless average exploitation in future years is higher than 1994-96 levels.

Relative abundance of northern squawfish declined slightly from 1995 in Bonneville Reservoir, Lower Monumental Dam tailrace, and Little Goose Dam tailrace. Mean abundance for 1994-96 was 48-60% of 1990-93 levels among areas sampled at least five years. Indices of consumption were lower than 1995 in all areas except during summer in the tailrace boat-restricted zones of Bonneville and The Dalles dams. Predation indices have declined 69% from pre- 1994 levels.

Decreases in proportional stock density were greater than could be explained by fluctuations in year-class strength, strongly suggesting that sustained removals may be altering the size structure of predator-sized northern squawfish. We found no evidence that northern squawfish have compensated in growth or fecundity in response to sustained exploitation.

INTRODUCTION

The goal of the Columbia River Northern Squawfish Management Program (NSMP) is to reduce mainstem mortality of juvenile salmonids attributed to predation by northern squawfish *Ptychocheilus oregonensis*. We established baseline levels of predation and described northern squawfish population characteristics prior to the implementation of sustained predator control fisheries by estimating northern squawfish abundance, consumption, and predation in Columbia River reservoirs (1990; Vigg et al. 1990), Snake River reservoirs (1991; Ward et al. 1993) and the unimpounded lower Columbia River downstream from Bonneville Dam (1992; Parker et al. 1994). We sampled Columbia River impoundments again in 1993 to evaluate changes from 1990 (Zimmerman et al. 1995). From 1994-96, we sampled in areas where sufficient numbers of northern squawfish could be collected to compare changes in predation among years (Knutsen et al. 1995; Zimmerman et al. in press). In this report we describe our activities and findings for 1996, and wherever possible, evaluate changes from previous years.

Our objectives in 1996 were to (1) evaluate the relative efficiency of each northern squawfish fishery by comparing exploitation rate, size composition, and incidental catch, (2) estimate reductions in predation on juvenile salmonids since the implementation of the NSMP, and (3) evaluate changes through 1996 in relative abundance, size structure, growth, and fecundity of northern squawfish in lower Columbia and Snake river reservoirs and in the Columbia River downstream from Bonneville Dam.

METHODS

Fishery Evaluation and Loss Estimates

Field Procedures

Three northern squawfish fisheries were conducted in 1996. The sport-reward fishery was implemented by the Washington Department of Fish and Wildlife (WDFW) from April 29-September 29 throughout the lower Columbia and Snake rivers. The dam-angling fishery was implemented by the Columbia River Inter-Tribal Fish Commission (CRITFC), Confederated Tribes of the Warm Springs Reservation of Oregon (CTWS), Confederated Tribes of the Umatilla Indian Reservation, Confederated Tribes and Bands of the Yakama Indian Nation (YIN), and the Nez Perce Tribe (NPT) from June 5- September 4 at Bonneville, The Dalles, John Day, McNary, Ice Harbor, Lower Monumental, Little Goose, and Lower Granite dams. A site-specific gill-net fishery was implemented by CRITFC, YIN, NPT, and CTWS from March 14-June 30 downstream from Bonneville Dam and in Bonneville, The Dalles, John Day, and McNary reservoirs.

We tagged and released northern squawfish to estimate exploitation rates for each fishery. We used electrofishing boats and gill nets to collect northern squawfish from March 25-June 20. We allocated equal sampling effort in all river kilometers (**RKm**) from RKm 7 1 through Priest Rapids Dam tailrace (RKm 639) on the lower Columbia River, and on the Snake River from the mouth through RKm 229, excluding Ice Harbor Reservoir. Northern squawfish greater than 240 mm fork length were tagged with a serially-numbered spaghetti tag and given a secondary mark (left pelvic fin clip). Tags were recovered from each fishery from April 22-September 29.

Data Analysis

We used mark-and-recapture data to compare exploitation rates of northern squawfish> 250 mm fork length among fisheries and reservoirs in 1996. Weekly estimates of exploitation for each fishery were calculated by dividing the number of tagged northern squawfish recovered by the number of tagged fish at large and summed to yield total exploitation rates (Beamesderfer et al. 1987). We also adjusted exploitation estimates for tag loss (4.2%), determined by the recovery of secondary-marked fish without tags.

We calculated 95% confidence intervals for the number of tagged fish recovered each week. We calculated confidence intervals for variables distributed in a Poisson distribution from Ricker (1975) for weeks when tagging and fishing occurred simultaneously. After tagging was complete, we calculated confidence intervals using the formula

m
$$\pm 1.96 \sqrt{m/n}$$
 (if mn > 30)

where

n = the number of sampling periods (weeks) remaining, and m = the mean number of tagged fish recovered per week (Elliott 1977).

Using this method, bounds were identical for all remaining weeks. We summed estimates for each week to give overall confidence limits.

We calculated mean fork length and compared length-frequency distributions among fisheries. We also compared incidental catch among fisheries for 1996 by determining the percent of the total catch composed of fish other than northern squawfish ≥ 250 mm fork length (incidental catch).

We used the "Loss Estimate Spreadsheet Model" (Zimmerman et al. 1995) to estimate predation on juvenile salmonids relative to predation prior to implementation of the NSMP. The model incorporates age-specific exploitation rates on northern squawfish and resulting changes in age structure to estimate changes in predation. We used a lo-year "average" age structure (based on catch curves) for a pre-exploitation base, and assumed constant recruitment. Age-specific consumption was incorporated; however, potential changes in consumption, growth, and fecundity due to removals were not considered likely. The model therefore estimates changes in potential predation related directly to removals. This in effect allows us to estimate the effects of removals if all variables except exploitation were held constant.

We estimated the potential relative predation in 1997 based on observed exploitation rates, and the eventual minimum potential predation assuming continuing exploitation at mean 199 l-96 levels. Because inputs to the model included three potential relationships between age of northern squawfish and consumption, and three estimates of exploitation (point estimate plus confidence limits), we computed nine estimates of relative predation for each year. We report the maximum, median, and minimum estimates.

Biological Evaluation

Field Procedures

To evaluate changes in relative abundance and consumption for northern squawfish, we used boat electrofishing to collect biological data from April 29-August 9 in the following areas: downstream from Bonneville Dam (RKm 117-121, RKm 171-177, and RKm 178-183), Bonneville Dam tailrace, Bonneville Reservoir, The Dalles Reservoir forebay, John Day Dam tailrace, John Day Reservoir, Little Goose Dam tailrace, Lower Granite Dam tailrace, and upper Lower Granite Reservoir (RKm 221-229). Sampling methods and gear specifications were as described in previous reports (Vigg et al. 1990; Ward et al. 1993; Parker et al. 1994; Zimmerman et al. 1995; Knutsen et al. 1995; Zimmerman et al. in press). Digestive tracts from all northern squawfish ≥ 250 mm fork length were preserved using methods described by Ward et al. (1995).

We collected biological data from all northern squawfish collected by electrofishing and from subsamples of fish harvested by the sport-reward and dam-angling fisheries to evaluate changes in population structure, growth, and reproduction. We determined fork length (mm), total body weight (g), sex (male, female or undetermined), and maturity (undeveloped or immature,

developing, ripe, or spent), and collected gonad (ripe females only) and scale samples. Fork length data for northern squawfish harvested by the site-specific gillnet fishery were provided by CRITFC.

Laboratory Procedures

We examined digestive tract contents of northern squawfish collected by electrofishing to measure relative consumption rates of juvenile salmonids by northern squawfish. Details of laboratory methods are given in Ward et al. (1995). We used scale samples from northern squawfish collected primarily by electrofishing for age determinations. We used gravimetric quantification (Bagenal 1968) to estimate fecundity of northern squawfish. Details of aging and fecundity procedures are given in Parker et al. (1995).

Data Analysis

We used catch per unit effort of standardized electrofishing runs as an index of northern squawfish density and calculated indices of northern squawfish abundance as the product of the northern squawfish density index and reservoir or area-specific surface area (Ward et al. 1995). We compared density and abundance indices from 1990-96 for all sampling areas.

The following formula was developed as a consumption index (CI) by Ward et al. (1995):

CI = 0.0209
$$T^{1.60}$$
 . $MW^{0.27}$. (s $GW^{-0.61}$)

where

 $T = water temperature (^{O}C),$

MW = mean predator weight (g),

S = mean number of salmonids per predator, and

GW = mean gut weight (g) per predator.

The consumption index is not a rigorous estimate of the number of juvenile salmonids eaten per day by an average northern squawfish; however, it is linearly related to the consumption rate of northern squawfish (Ward et al. 1995). Spring (May-June) and summer (July-September) consumption indices were compared from 1990-96 for all sampling areas. We used the product of abundance and consumption indices to calculate predation indices for spring and summer periods. We compared predation indices among years for reservoirs and areas where data had been collected each year. We plotted the daily juvenile salmonid passage index at lower Columbia and Snake River dams to compare timing of consumption index sampling with concentrations of juvenile salmonids present in each area.

Because fishery exploitation rates increase with size of northern squawfish (Zimmerman et al. 1995) sustained fisheries should decrease the abundance of large fish relative to the abundance of smaller fish. We used proportional stock density (PSD; Anderson 1980) where PSD = 100 *(number of fish at least quality length / number of fish at least stock length) to compare size structure of northern squawfish populations among years in the Columbia River downstream from

Bonneville Dam, Bonneville Reservoir, and John Day Reservoir. Stock and quality sizes for northern squawfish have been defined as 250 and 380 mm fork length, respectively (Beamesderfer and Rieman 1988; Parker et al. 1995).

Comparisons of PSDs among years may be biased by fluctuating year-class strengths that influence the number of stock-size fish (Mesa et al. 1990) and size-selectivity of sampling gear (Beamesderfer and Rieman 1988). To reduce bias, we used information on relative year-class strengths and natural mortality rates of northern squawfish to estimate PSDs that would be expected with and without program implementation and used size-selectivity of our sampling gear to adjust observed PSD estimates (Knutsen et al. 1995). We then compared observed and expected PSDs.

To evaluate changes in growth rate after implementation of the NSMP, we used observed length-at-age data for female northern squawfish from the Columbia River downstream from Bonneville Dam and Bonneville, John Day, and Lower Granite reservoirs. We determined regression parameters (slope and y-intercept) for fork length on age using only those ages where growth rate was linear (ages 5 through 14 downstream from Bonneville Dam and Bonneville Reservoir, ages 5 through 13 in John Day Reservoir, and ages 6 through 11 in Lower Granite Reservoir). We compared growth among years using joint 95% family confidence regions for estimates of parameter pairs (Neter et al. 1985). Parameter pairs were considered significantly different if point estimates (center-point of ellipse) were outside the confidence region for another year.

We calculated mean fecundity (number of developed eggs per female) and mean relative fecundity (number of developed eggs per gram of body weight) for northern squawfish from the Columbia River downstream from Bonneville Dam, and Bonneville, John Day, and Lower Granite reservoirs. We determined parameters for the regression of fecundity on fork length and compared relationships among years for each area using joint 95% family confidence regions for estimates of parameter pairs (Neter et al. 1985).

RESULTS

Fishery Evaluation and Loss Estimates

We tagged and released 1,430 northern squawfish throughout the lower Columbia and Snake rivers in 1996. A total of 162 tagged fish were recaptured in the three fisheries: 152 in the sport-reward fishery, four in the dam-angling fishery, and six in the site-specific gill-net fishery.

Total exploitation of northern squawfish in 1996 was 12.9%, and ranged from 0.0% in Lower Monumental Reservoir to 18.2% in McNary Reservoir (Figure 1; **Appendix A**). Exploitation of northern squawfish by the sport-reward fishery was 12.1% in 1996. Sport-reward exploitation increased from 1995 in Bonneville, The Dalles, Little Goose, and Lower Granite reservoirs, decreased downstream from Bonneville Dam, and did not change in John Day and Lower

Monumental reservoirs (Figure 1; **Appendix Table A-l).** Dam-angling exploitation was 0.3%, and 1996 exploitation rates declined in all areas except John Day Reservoir (Figure 1; **Appendix Table A-2).** Exploitation by the site-specific gill-net fishery totaled 0.5% and declined or remained the same in all areas (Figure 1; **Appendix Table A-3).** A sufficient number of tagged fish were captured in 1996 (mn > 30) to calculate 95% confidence intervals for exploitation estimates (all fisheries combined) downstream from Bonneville Dam and systemwide.

Mean fork length of northern squawfish harvested systemwide in 1996 was 355 mm in the sport-reward fishery, 391 mm in the dam-angling fishery, and 408 mm in the site-specific fishery (Figure 2). Mean fork length increased from 1995 in the sport-reward (8.6%) and dam-angling (6.5%) fisheries, and decreased slightly (0.7%) in the site-specific fishery.

Mean sizes of northern squawfish harvested in specific areas by the sport-reward fishery in 1996 were within the range for other years, except fish harvested downstream from Bonneville Dam and in Lower Granite Reservoir were larger than in any other year (Table 1). Mean fork length of northern squawfish harvested by the dam-angling fishery in 1996 was higher than 1995 in Bonneville Dam tailrace and John Day Reservoir, and lower than any other year in Bonneville Reservoir. Sample size of northern squawfish harvested by the dam-angling fishery in The Dalles Reservoir was inadequate to estimate mean fork length. Mean fork length of northern squawfish harvested by the gill-net fishery in 1996 was higher than 1995 downstream from Bonneville Dam and in John Day Reservoir, lower in Bonneville and McNary reservoirs, and similar to 1995 in the Dalles Reservoir, Sample sizes were small (<200) for all locations except Bonneville Reservoir.

In 1996, the various fisheries reported 65,814 incidentally-caught fish including northern squawfish <250 mm fork length (Table 2). The incidental catch rate was 27.0% in the sport-reward fishery, 3.6% in the dam-angling fishery, and 54.6% in the gill-net fishery. Northern squawfish <250 mm, other cyprinids, smallmouth bass, catostomids, and white sturgeon were the most common incidentally-caught fish. Salmonids made up only 0.3% of the total catch and 1.2% of the incidental catch for all fisheries combined. The proportion of predator-sized (≥250 mm fork length) northern squawfish relative to the total number of northern squawfish harvested was highest in the dam-angling fishery (98.7%) and lowest in the sport-reward fishery (87.1%).

Results from the "Loss Estimate Spreadsheet Model" indicate that potential predation by northern squawfish on juvenile salmonids in 1997 may range from 45% to 75% of pre-program levels, with a median estimate of 62% (Figure 3). Continued exploitation at mean 1991-96 levels will not result in further reductions in predation; predation will remain similar to 1997 levels unless exploitation is increased.

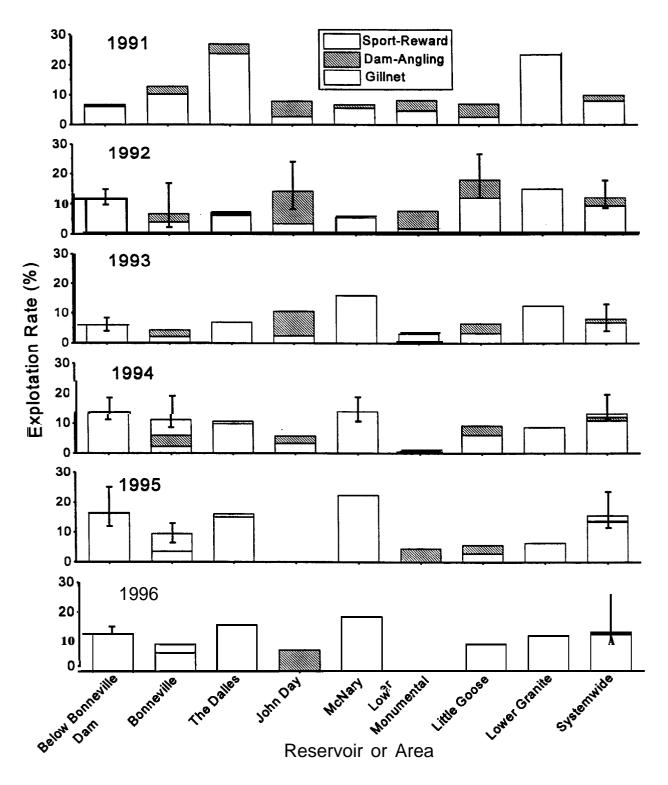


Figure 1. Exploitation of northern squawfish \geq 250 mm fork length by fishery and area, 1991-96. Confidence intervals for all fisheries combined (96%; vertical bars) are included where a sufficient number of tagged northern squawfish were recaptured to estimate upper and lower bounds.

Table 1. Mean fork length (mm) of northern squawfish harvested from 1990-96 in the lower Columbia and Snake rivers by the sport-reward, dam-angling, and site-specific gill-net fisheries.

Mean fork length (mm)

Fishery: Location	1990	1991	1992	1993	1994	1995	1996
Dam Angling:							
Bonneville Dam Tailrace	414	417	388	390	376	365	370
Bonneville Reservoir	407	417	416	415	413	366	363
The Dalles Reservoir	421	404	380	420	390	343	
John Day Reservoir	416	414	417	416	437	389	418
McNary Reservoir	393	393	375	408			
Ice Harbor Reservoir		375	369	414			
Lower Monumental Reservoir		325	309	341			
Little Goose Reservoir		380	346	373	370		
Lower Granite Reservoir				377			
Sport Reward:							
Downstream from Bonneville Dam		332	337	316	337	325	345
Bonneville Reservoir		343	347	312	323	305	344
The Dalles Reservoir		344	369	369	358	359	349
John Day Reservoir	377	370	367	370	329		
McNary Reservoir		354	356	358	366	323	359
Ice Harbor Reservoir		357	360	317	407		
Lower Monumental Reservoir		338	330	307	428		
Little Goose Reservoir		312	347	344	376		
Lower Granite Reservoir		343	345	362	348	322	375
Gill Net:							
Downstream from Bonneville Dam						374	391
Bonneville Reservoir				371	411	416	409
The Dalles Reservoir					395	396	397
John Day Reservoir					366	370	403
McNary Reservoir					387	363	353
Ice Harbor Reservoir						386	
Lower Monumental Reservoir						345	
Little Goose Reservoir						379	
Lower Granite Reservoir					377	370	

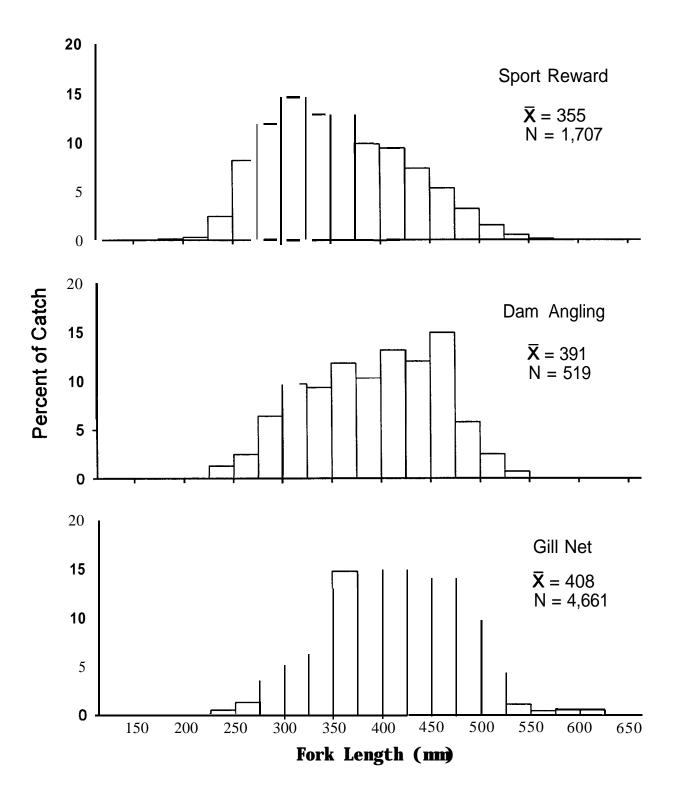


Figure 2. Size composition and mean fork length of northern squawfish from subsamples of fish harvested systemwide by sport-reward, dam-angling, and site-specific gill-net fisheries during 1996. N = subsample size.

Biological Evaluation

Density indices for northern squawfish in 1996 were generally similar to those of 1995 (**Appendix Table B-I).** Total (all sampling areas combined) density increased slightly downstream from Bonneville Dam, in John Day Reservoir, and in the upper portion of Lower Granite Reservoir, and decreased in Bonneville, The Dalles, Lower Monumental, and Little Goose Reservoirs. Density indices for seven of the 19 areas sampled were the lowest recorded since biological evaluation began in 1990.

Relative abundance of northern squawfish declined slightly in Bonneville Reservoir, and in the tailrace zones of Lower Monumental and Little Goose reservoirs in 1996 relative to previous years (Figure 4; **Appendix Table B-2).** Abundance in the remaining areas were similar to or slightly higher than abundance in 1994-95. Mean abundance in Bonneville Dam tailrace, and Bonneville, The Dalles, and John Day reservoirs (areas sampled at least five years) was 48-60% lower from 1994-96 than 1990-93.

Consumption in spring was lower in 1996 relative to previous years in all locations (**Appendix Table B-3**). Consumption in summer was also lower than previous years except in the boat restricted zones of Bonneville Dam tailrace and The Dalles Reservoir tailrace (**Appendix Table B-4**). We found no juvenile salmonids in the digestive tracts of nor-them squawfish collected at seven of 14 summer sampling sites, which resulted in consumption indices of zero for these sites. Among areas sampled at least five years, mean consumption from 1994-96 was 2 1-4 1% lower in spring and 13-64% lower in summer than 1990-93.

Predation indices in spring 1996 were lower than all prior years in every location (Figure 5; **Appendix Table B-5).** Predation in summer 1996 was higher than predation in 1994-95 in most tailrace and forebay areas, but was generally lower than predation prior to 1994 (Figure 6; **Appendix Table B-6).** The mean decline in predation from pre- to post-1994 levels ranged from 40% downstream from Bonneville Dam to 86% in Little Goose Dam tailrace. The mean decline in all areas was 69%. Timing of consumption/predation sampling coincided with peaks in downstream passage of juvenile salmonids in all sampling areas during spring, and at John Day Dam (forebay and tailrace), McNary Dam (tailrace), and McNary Reservoir (mid-reservoir) during summer (**Appendix C**).

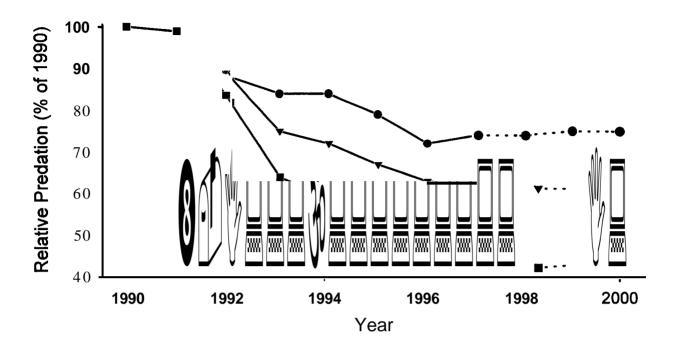


Figure 3. Maximum (circles), median (triangles), and minimum (squares) estimates of potential predation on juvenile salmonids by northern squawfish relative to predation prior to implementation of the northern squawfish management program. Dashed lines indicate predation in future years if exploitation is maintained at mean 1991-96 levels.

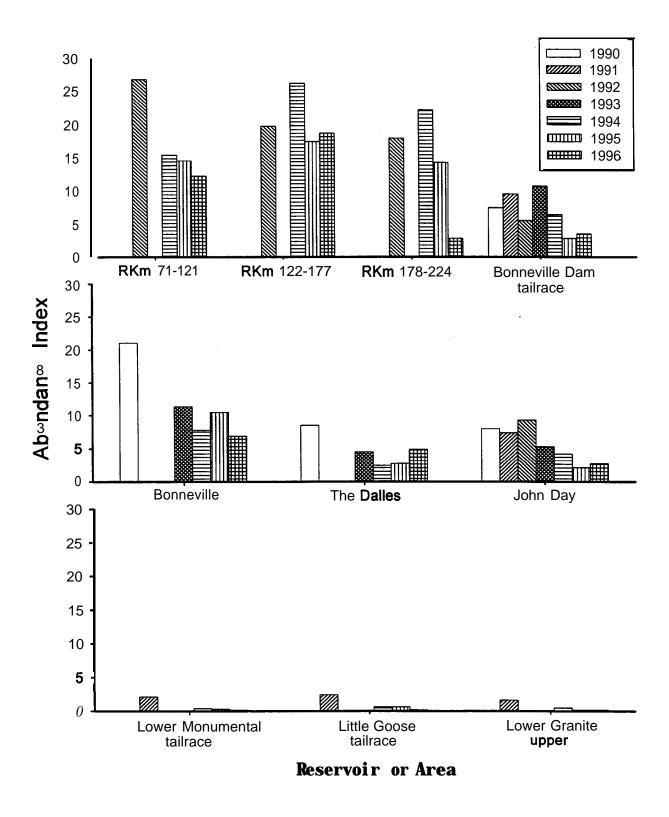


Figure 4. Index of northern squawfish abundance from 1990-96 for sampling locations in the lower Columbia and Snake rivers. RKm = river kilometer.

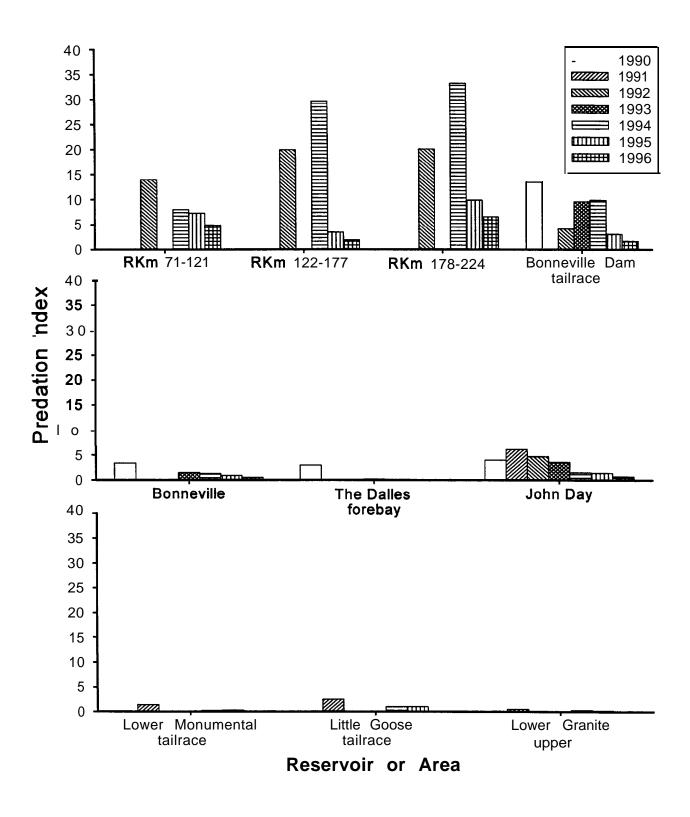


Figure 5. Index of predation by northern squawfish during spring from 1990-96 for sampling locations in the lower Columbia and Snake rivers. RKm = river kilometer.

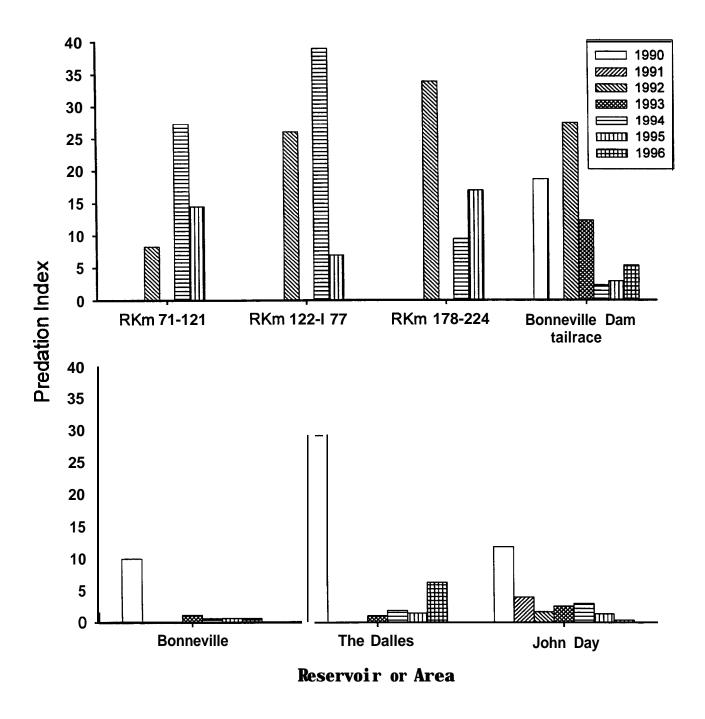


Figure 6. Index of predation by northern squawfish during summer from 1990-96 for sampling locations in the lower Columbia and Snake rivers. Predation indices for The Dalles Reservoir excludes the mid-reservoir. R.Km = river kilometer.

Year-class strength of northern squawfish was highly variable; however, variations were generally similar among areas (Figure 7). Differences among areas were less than 25% except in 1987. Year classes were generally stronger from 1989-92 than in 1987 and 1988.

Proportional stock density of northern squawfish has generally decreased since implementation of the NSMP (Figure 8). Observed PSDs initially remained stable or increased, as a relatively strong 1985 year class was recruited from stock to quality size. Observed PSDs then generally decreased as relatively strong 1989-91 year classes were recruited to stock size. In Bonneville and John Day reservoirs, observed PSDs were lower than PSDs expected without implementation of the management program. After 1993, observed PSDs were very similar to PSDs expected with implementation of the NSMP. Observed PSDs followed expected trends downstream from Bonneville Dam except in 1995, when the observed PSD was higher than expected.

Estimates of mean fecundity and mean relative fecundity in each reservoir or reach increased from 1995 to 1996; however, fecundity in 1996 was generally similar to or less than fecundity from 199 l-94 (Table 3). Although regression parameters for fecundity-fork length relationships differed among years, there was no evidence that fecundity of northern squawfish increased from 199 l-96 (Table 4).

Regression parameters for length-age relationships differed among some years for female northern squawfish downstream from Bonneville Dam, and in Bonneville and John Day Reservoirs, but were similar among years in Lower Granite Reservoir (Table 5). There was no trend in increasing growth from 1990-96 in any area, despite differences in growth among a few years.

DISCUSSION

Rieman and Beamesderfer (1990) predicted that sustained exploitation of northern squawfish > 275 mm fork length at an annual rate of 10-20% would reduce losses of juvenile salmonids to predation by 50%. Total systemwide exploitation by the three fisheries in 1996 was 12.9%, the third consecutive year that exploitation has exceeded 10%. Systemwide exploitation from 199 1-95 averaged 11.7% (Zimmerman et al. in press). As in previous years, sport-reward exploitation greatly exceeded exploitation by other fisheries. The dam-angling and site-specific gill-net fisheries, while contributing less to exploitation, harvested localized concentrations of northern squawfish which may have aggregated to feed on juvenile salmonids (Beamesderfer and Rieman 1991; Poe et al. 1991; Collis et al. 1995). These fisheries also harvested larger northern squawfish than the sport-reward fishery, and captured a much smaller number of other species.

If exploitation rates remain similar to mean 1991-96 levels, it is likely that no further reductions in potential predation will be realized. Predation will remain at approximately 62% of pre-program levels. Exploitation rates lower than mean 199 l-96 levels will result in increases in

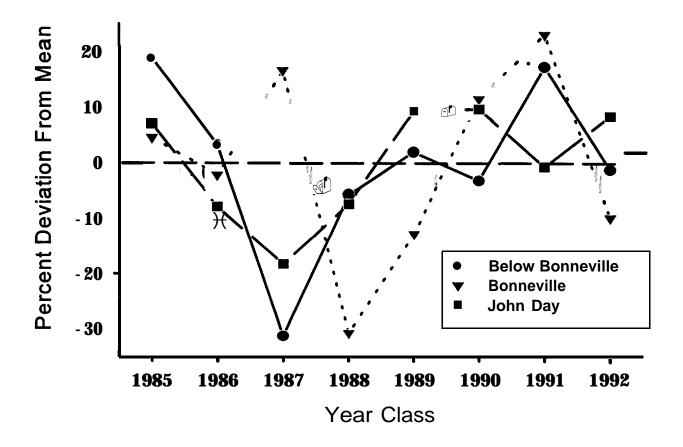


Figure 7. Index of relative year-class strength of northern squawfish in the Columbia River downstream from Bonneville Dam, in Bonneville Reservoir, and in John Day Reservoir.

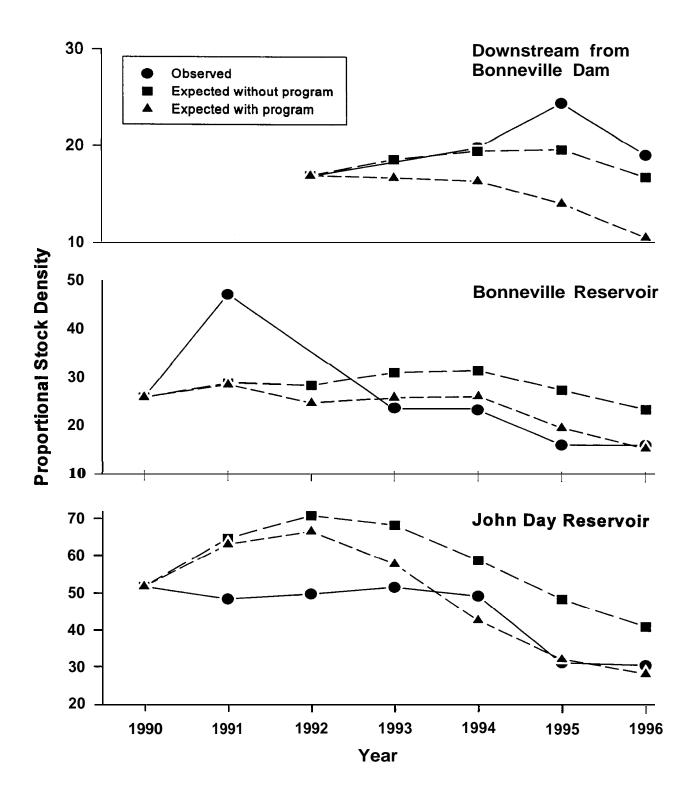


Figure 8. Observed and expected proportional stock density with and without implementation of the Northern Squawfish Management Program from 1990-96 in the Columbia River downstream from Bonneville Dam, in Bonneville Reservoir, and in John Day Reservoir.

Table 3. Mean fecundity (number of developed eggs per female), mean relative fecundity (MRF; number of developed eggs per gram of body weight), and sample size (N) of northern squawfish in selected areas of the Columbia and Snake rivers, 1991-96.

Location, parameter	1991	1992	1993	1994	1995	1996
Downstream from						
Bonneville Dam						
Mean fecundity	37,500	25,069	22,410	26,717	18,865	24,047
MRF	40.01	38.74	37.68	37.37	31.32	40.79
N	59	247	267	84	121	39
Bonneville Reservoir						
Mean fecundity	31,225	33,640	30,444	29,3 13	18,550	26,413
MRF	45.17	35.96	32.15	32.43	22.27	33.90
N	37	105	101	100	6	47
John Day Reservoir						
Mean fecundity	30,699	32,026	26,016	27,63 8	17,792	36,487
MRF	28.90	32.40	24.72	24.93	20.80	31.05
N	66	109	95	60	12	51
Lower Granite Reservoir						
Mean fecundity	23,251	24,851	32,472	21,926	23,991	25,795
MRF	30.29	27.51	27.79	34.28	21.80	29.23
N	44	30	17	5	5	65

Table 4. Differences among years in joint 95% family confidence regions for parameter pairs of the regression of fecundity on fork length (mm) of northern squawfish. * = significant difference between years; ns = no difference between years.

		St	atistical	Significa	nce		Par	ameters
Location,								
Year	1991	1992	1993	1994	1995	1996	Slope	Intercept
Downstream	from							
Bonneville Da	ım							
1991							254.6	-66,804
1992	*						106.5	-15,081
1993	*	ns					127.1	-23,421
1994	*	*	*				180.7	-41,757
1995	*	*	*	*			67.6	-6,037
1996	*	ns	*	ns	*		172.3	-36,648
Bonneville R	eservoir							
1991							229.7	-58,353
1992	ns						145.5	-26,444
1993	*	ns					105.4	-13,113
1994	*	*	ns				98.5	-11,666
1996	ns	ns	*	*			172.0	-40,410
John Day Res	servoir							
1991							110.8	-19,152
1992	*						114.0	-16,682
1993	ns	*					115.5	-24,036
1994	ns	ns	ns				146.8	-36,626
1995	ns	*	ns	ns			96.0	-21,367
1996	*	ns	*	*	*		196.9	-51,113
Lower Granit	e							
Reservoir								
1991							152.7	-34,631
1992	ns						161.8	-40,845
1996	ns	ns					133.6	-27,3 57

Table 5. Differences among years in joint 95% family confidence regions for parameter pairs of the regression of fork length (mm) on age of female northern squawfish. * = significant difference between years; ns = no difference between years.

		Statistical Significance				Parameters			
Location,									
Year	1990	1991	1992	1993	1994	1995	1996	Slope	Intercept
Downstream from									
Bonneville Dam									
1990								21.2	195.9
1991	*							28.8	106.5
1992	*	*						28.8	131.5
1993	*	ns	ns					28.4	122.1
1994	ns	*	ns	*				25.3	187.1
1995	ns	ns	ns	ns	ns			28.8	131.1
1996	*	*	ns	ns	ns	ns		26.0	152.1
Bonneville Reservoir									
1990								21.7	188.0
1993	*							28.8	127.4
1994	*			*				27.6	168.9
1995	ns			*	*			24.0	183.2
1996	ns			*	*	ns		22.6	192.2
John Day Reservoir									
1990								20.0	206.3
1991	ns							22.4	193.0
1992	*	ns						28.6	136.2
1993	*	ns	ns					26.4	158.6
1994	ns	ns	*	ns				21.5	208.8
1995	ns	ns	ns	ns	ns			24.6	171.6
1996	ns	ns	ns	ns	ns	ns		23.0	192.7
Lower Granite									
Reservoir									
1991								32.5	92.3
1993		ns						29.6	108.7
1995		ns		ns				26.8	149.4
1996		ns		ns		ns		31.3	124.7
1//0		110		110		110		31.5	

potential predation. Because exploitation from 1994-96 was higher than exploitation from 199 l-93, further reductions in predation may be feasible. These reductions will be small (1% to 4%) unless average exploitation is higher than 1994-96 levels.

Reducing the number of large northern squawfish may improve salmonid survival if remaining northern squawfish do not consume salmonids at a higher rate (Beamesderfer et al. 1990). Declining abundance and consumption of northern squawfish contributed to declines in predation on juvenile salmonids from 1990-96 in all areas. Our results indicate that northern squawfish are not consuming more juvenile salmonids in response to decreased abundance of northern squawfish. Density (and therefore abundance and predation) indices for the tailrace boat-restricted zones of The Dalles, Lower Monumental, and Little Goose reservoirs may be biased low because river conditions during 1996 limited our sampling of these areas.

Decreases in PSD were greater than could be explained by fluctuations in year-class strength, and indicate that sustained removals are changing the size structure of predator-sized northern squawfish. Observed and expected PSD values often differed in the early years of the NSMP, in part because estimates of expected PSD incorporate estimates of natural mortality and growth. Annual variations in mortality and growth are unpredictable; we therefore used estimates of natural mortality developed prior to sustained removals, and pooled 1990-96 growth data to estimate age-specific lengths. These data should be representative of long-term averages, but will not reflect annual variation around those averages. Differences between observed and expected PSD values have become smaller over a period of years, as annual variations in mortality and growth "average out."

If northern squawfish compensate for sustained exploitation with increased growth or fecundity rates, relative benefits of the NSMP will be diminished (Beamesderfer et al. 1996). While our estimates of fecundity and growth of female northern squawfish differed in some years, there was no evidence that northern squawfish are compensating for decreased abundance by increasing consumption and consequently growing faster, or by increasing fecundity.

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APPENDIX A

Exploitation of Northern Squawfish by Reservoir and Fishery, 199 1-96

Appendix Table A-l. Exploitation rates (%) of northern squawfish2 250 mm fork length for the sport-reward fishery, 199 l-96.

Systemwide	8.5	9.3	6.8	10.9	13.4	12.1
Lower Granite	20.0	15.0	12.5	8.7	6.4	11.7
Little Goose	2.4	11.9	3.3	6.1	2.9	8.9
Lower Monumental	4.5	1.8	3.1	0.8	0.0^a	0.0^{a}
Ice Harbor	1.0	b	b	b	b	b
McNary	5.3	5.6	15.9	14.0	22.4	18.2
John Day	2.8	3.4	2.4	3.2	0.0"	0.0"
The Dalles	23.6	6.2	7.0	9.8	14.9	15.5
Bonneville	10.9	4.0	2.1	2.2	3.5	6.1
Bonneville Dam	7.6	11.4	6.0	13.6	16.1	12.7
Downstream from						
Reservoir	1991	1992	1993	1994	1995	1996
Area or						

<sup>a Northern squawfish harvested, but no tags recovered.
b No northern squawfish tagged.</sup>

Appendix Table A-2. Exploitation rates (%) of northern squawfish> 250 mm fork length for the dam-angling fishery, 199 1-96.

Area or Reservoir	1991	1992	1993	1994	1995	1996
Downstream from						
Bonneville Dam	0.9	0.2	0.0"	0.1	0.2	0.0"
Bonneville	2.7	2.8	2.2	3.7	0.0"	0.0"
The Dalles	3.3	1.0	0.0"	0.0"	0.0"	0.0"
John Day	5.2	10.8	8.1	2.6	0.0"	7.0
McNary	1.1	0.0"	0.1	0.0"	0.0"	0.0"
Ice Harbor	3.4	b	b	b	b	b
Lower Monumental	3.1	5.9	0.0^{a}	0.0"	4.5	0.0^a
Little Goose	4.2	6.0	3.3	3.1	2.8	0.0^a
Lower Granite	< 0.1	0.0"	0.0"	0.0"	0.0"	0.0^a
Systemwide	2.2	2.7	1.3	1.1	0.3	0.3

a Northern squawfish harvested, but no tags recovered.
 b No northern squawfish tagged.

Appendix Table A-3. Exploitation rates (%) of northern squawfish> 250 mm fork length for the site-specific gill-net fishery, 1994-96.

Area or Reservoir	1994	1995	1996
Downstream from Bonneville Dam	a	0.2	0.0^{b}
Bonneville	5.3	5.9	3.0
The Dalles	0.9	1.1	0.0^{b}
John Day	0.0^{b}	0.0^{b}	0.0^{b}
McNary	0.0^{b}	0.0^{b}	0.0^{b}
Ice Harbor	ac	c	ac
Lower Monumental	0.0^{b}	0.0^{b}	a
Little Goose	a	0.0^{b}	a
Lower Granite	0.0^{b}	0.0^{b}	a
Systemwide	1.2	1.9	0.5

Appendix Table A-4. Total exploitation rates of northern squawfish2 250 mm fork length, 199 l-96.

Area or reservoir	1991	1992	1993	1994	1995	1996
Downstream from						
Bonneville Dam	8.6	11.7	6.0	13.8	16.5	12.7
Bonneville	13.6	6.8	4.3	11.2	9.4	9.1
The Dalles	26.9	7.2	7.0	10.7	16.0	15.5
John Day	8.0	14.2	10.5	5.8	0.0	7.0
McNary	6.5	5.6	16.0	14.0	22.4	18.2
Ice Harbor	4.4	a	a	a	a	a
Lower Monumental	7.6	7.7	3.1	0.8	4.5	0.0
Little Goose	6.6	17.9	6.6	9.2	5.7	8.9
Lower Granite	20.0	15.0	12.5	8.7	6.4	11.7
Systemwide	10.7	12.0	8.1	13.2	15.5	12.9

^a No northern squawfish tagged.

^a No fishing effort.
^b Northern squawfish harvested, but no tags recovered.
^c No northern squawfish tagged.

Appendix Table A-5. Dates for each sampling period in 1996.

Period	Dates	Period	Dates
1	Before April 1	15	July 1 - July 7
2 3	April 1 - April 7	16	July 8 - July 14
3	April 8 - April 14	17	July 15 - July 2 1
4	April 15 -April 21	18	July 22 - July 28
5	April 22 - April 28	19	July 29 - August 4
6	April 29 - May 5	20	August 5 - August 11
7	May 6 - May 12	21	August 12 - August 18
8	May 13 - May 19	22	August 19 - August 25
9	May 20 - May 26	23	August 26 - September 1
10	May 27 - June 2	24	September 2 - September 8
11	June 3 - June 9	25	September 9 - September 15
12	June 10 - June 16	26	September 16 - September 22
13	June 17 - June 23	27	September 23 - September 29
14	June 24 - June 30		-

Appendix Table A-6. Exploitation of northern squawfish downstream from Bonneville Dam in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

m:		I	Recaptures	S		Exploitation			
Time neriod	T	sport	Dam	Net	Misc.	M	sport	Dam	Net
1									
2	83								
2 3	334					83			
4	358					417			
5						775			
6		3				775	0.0039		
7		2				772	0.0026		
8		4				770	0.0052		
9		2	-			766	0.0026		
10		2				764	0.0026		
11		1				762	0.0013		
12		4				761	0.0053		
13		2			1	757	0.0026		
14		9				754	0.0119		
15		11				745	0.0148		
16		11				734	0.0150		
17		8				723	0.0111		
18		7				715	0.0098		
19		7			2	708	0.0099		
20		3				699	0.0043		
21		3				696	0.0043		
22		1				693	0.0014		
23		1				692	0.0014		
24		3				691	0.0043		
25		2				688	0.0029		
26		1				686	0.0015		
27		2				685	0.0029		
Γotal	775	89	0	0	3		0.1216	0.0000	0.0000
Adjusted	for tag lo	SS					0.1268	0.0000	0.0000

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Appendix Table A-7. Exploitation of northern squawfish in Bonneville Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

m:		R	ecapture	S			Ex	ploitation	ı
Time Period	Т	Sport	Dam	Net	Misc.	M	Sport 1	O a m	Net
1	28								
2	187					28			
3						215			
4					••	215			
5				2	÷	215	\$60 APA		0.0093
6		1		1		213	0.0047		0.0047
7						211			
8		1				211	0.0047		
9						210			
10		1				210	0.0048		
11		1		1		209	0.0048		0.0048
12				2	1	207			0.0097
13		2				204	0.0098		
14		1				202	0.0050		
15						201			
16		3				201	0.0149	***	
17					2	198			
18						196			
19						196			
20						196			
21					1	196			
22		2				195	0.0103		er er
23						193			
24						193			
25						193			
26						193			
27						193			
Total	215	12	0	6	4		0.0589	0.0000	0.0284
Adjusted	for tag lo	oss				0.0614	0.0	000 0.	0296

Appendix Table A-S. Exploitation of northern squawfish in The Dalles Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

Time period	Recaptures						Exploitation		
	Т	Sport	Dam	Net	Misc.	M	Sport	Dam	Net
1	26								
2						26			
2 3						26			
4						26			
5						26			
6						26			-
7	46					26			
8		1				72	0.0139		
9					==	71			
10		1				71	0.0141		
11		1				70	0.0143		
12						69			
13						69			
14		2				69	0.0290		
15		1				67	0.0149		
16		2				66	0.0303		
17						64			
18					1	64			
19		1				63	0.0159		
20		1				62	0.0161		
21						61			
22						61			
23						61			
24						61			
25						61			
26						61			
27						61			
Total	72 1	0 0	0	1		0.1485	0.0000		0.0000
Adjusted for tag loss						0.1547	0.0000		0.0000

Appendix Table A-9. Exploitation of northern squawfish in John Day Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

T:		Recap	tures			Exp	loitation
Time period	T	Sport	Dam	Misc.	M	sport	Dam
1	23						
2					23		==
3					23		
4					23		
5					23		
6 7					23		
7	3			1	23		
8	19			1	25		
9	19			1	43	***	en 10
10					61		==
11					61		==
12				1	61		
13			1		60	00	0.0167
14					59		
15					59		
16					59	100 etc.	
17					59		•=
18			3		59	00	0.0508
19					56	***	
20					56		
21				1	56		
22					55		
23					55		* *
24					55		
25					55		
26					55		
27					55		
Total	64	0	4	5		0.0000	0.0675
Adjuste	ed for tag	g loss				0.0000	0.0703

Appendix Table A-10. Exploitation of northern squawfish in McNary Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

Time		Reca	ptures			Exp	oloitation
period	Т	Sport	Dam	Misc.	M	sport	Dam
1	4				dispos della		***
					4	10 4	
2 3					4		
4					4	100 444	
4 5	104				4	W 20	
6		1	~~		108	0.0093	
7		1			107	0.0093	
8					106		
9	7	1			106	0.0094	
10	22				112		
11	18				134		
12					152		
13					152		
14					152		
15		1			152	0.0066	
16		2			151	0.0132	
17		4			149	0.0268	
18		6			145	0.0414	
19		1			139	0.0072	
20		1			138	0.0072	
21					137	-	
22		2			137	0.0146	
23					135		
24		1			135	0.0074	
25		1			134	0.0075	
26		2			133	0.0150	
27					131		
Total	155	24	0	0		0.1750	0.0000
Adjuste	d for tag l	OSS				0.1824	0.0000

Appendix Table A-l 1. Exploitation of northern squawfish in Lower Monumental Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

Time		Reca	ptures			Explo	oitation
period	Т	sport	Dam	Misc.	M	sport	Dam
1							
2						NO 100	
3							
2 3 4 5							
5						NO 45	
6 7							==
7							
8 9							
9							
10							
11	50						
12					50		
13					50		
14					50		
15	100 440				50		
16					50		
17					50		
18					50		
19					50		
20					50		
21					50		
22					50		
23					50		
24					50		
25					50		***
26					50		
27					50		
Total	50	0	0	0		0.0000	0.0000
Adiusted	for tag	loss				0.0000	0.0000

Appendix Table A-12. Exploitation of northern squawfish in Little Goose Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

Time		Reca	ptures			Exploitation		
period	T	Sport	Dam	Misc.	M	sport	Dam	
1								
3								
4								
2 3 4 5								
				₩=				
6 7								
8								
9								
10						••		
11								
12								
13	24	min ain						
14					24		==	
15					24			
16					24			
17					24			
18		1			24	0.0417		
19					23		<u>~~~</u>	
20					23			
21		-			23		***	
22		1			23	0.0435		
23					22			
24					22			
25					22			
26					22			
27					22			
Γotal	24	2	0	0		0.085 1	0.0000	
Adiusted	for tag	loss				0.0887	0.0000	

Appendix Table A-13. Exploitation of northern squawfish in Lower Granite Reservoir in 1996. T = number of fish tagged. M = number of tagged fish at large. Misc. = tagged fish recaptured outside the program area or fisheries.

Γime .		Reca	aptures			Exploitation		
period	T	Spor	t Dam	Misc.	M	sport	Dam	
1								
			-					
3								
2 3 4 5								
5								
6	75	1				0.0133		
7					74			
8					74			
9					74			
10					74			
11		1			74	0.0135		
12					73			
13		1			73	0.0137		
14		1			72	0.0139		
15					71			
16		2			71	0.0282		
17		1			69	0.0145		
18					68			
19					68			
20					68			
21					68			
22					68			
23					68			
24		1			68	0.0147		
25					67			
26					67			
27					67			
Total	75	8	0	0		0.1118	0.0000	
Adjusted	for tag	loss				0.1165	0.0000	

Appendix Table A-14. Exploitation of northern squawfish systemwide in 1996. T = number of fish tagged. M = number of tagged **fish** at large. Misc. = tagged fish recaptured outside the program area or fisheries.

T:			Recaptur	es			Ex	ploitation	
Time period	T	Sport	t Dam	Net	Misc.	M	Sport	Dam	Net
1	81					***			
2	270					81			
3	334					351			
4	358					685			
5	104			2		1043			0.0019
6	75	6		1		1145	0.0052		0.0009
7	49	3			1	1213	0.0025		
8	19	6			1	1258	0.0048		
9	26	4				1270	0.003 1	~	
10	22	4				1292	0.003 1		
11	68	4		1		1310	0.003 1		0.0008
12		5		2		1373	0.0036		0.0015
13	24	6	1			1366	0.0044	0.000	7
14		13			***	1383	0.0094		
15	**	13			ee ep	1370	0.0095		
16		20			1	1357	0.0147		
17		14			1	1336	0.0105		
18		15	3			1321	0.0114	0.0023	
19		9			2	1303	0.0069		
20		5				1292	0.0039		
21		5			-	1287	0.0039		
22		6				1282	0.0047		
23		1				1276	0.0008		
24		5				1275	0.0039		
25		3				1270	0.0024		
26		3				1267	0.0024		
27		2				1264	0.0016		
Total	1430	152	4	6	6		0.1156	0.0030	0.0050
Adjusted for	or tag loss	S					0.1205	0.003 1	0.0052

APPENDIX B
Density, Abundance, Consumption, and Predation Indices for Sampling Locations in the Lower Columbia and Snake Rivers, 1990-96

Appendix Table B-l. Indices of northern squawfish density from 1990-96 for sampling areas in the lower Columbia and Snake rivers. RKm = river kilometer, BRZ = boat restricted zone.

Location, area	1990	1991	1992	1993	1994	1995	1996
Location, area	1990	1771	1772	1773	1777	1773	1770
Below							
Bonneville Dam							
RKm 71-121			1.3(68)		,	0.9(45)	` ′
RKm 122-177			1.6(65)			1.4(36)	
RKm 178-224			2.4(64)		,	3) 1.1 (4 0)	, ,
Tailrace	5.8(26)	6.9(27)	3.4(37)	9.6(16)	` '	2.3(16)	2.8(24)
Tailrace BRZ	13.7(13)	19.0(13)	12.9(23)	14.5(9)	18.9(8)	4.3(8)	5.8(7)
Bonneville Reservoir							
Forebay	5.7(47)			2.2(35)	2.4(97)	2.4(79)	1.3(80)
Mid-reservoir	2.1(52)			1.2(28)	0.7(84)	1.0(45)	0.7(57)
Tailrace	0.5(37)			1.1(25)	0.6(60)	1.1(80)	0.8(69)
Tailrace BRZ	5.5(15)			1.5(6)	6.8(8)		
The Dalles Reservoir							
Forebay	1.1(62)			1.2(3 1) 0.0	6(92) 0.6(62) 0.4(59
Tailrace	2.8(45)			0.7(21)	0.7(40)		
Tailrace BRZ	21.5(11)			10.8(5)	5.5(8)	3.5(8)	1.0(3)
John Day Reservoir							
Forebay	0.7(56)	0.7(61)	1.3(68)	0.6(44)	0.7(91)	0.3(75)	0.3(54)
Mid-reservoir	0.3(61)	0.2(58)	0.3(62)	0.2(43)	0.1(43)	0.1(94)	0.1(52)
Tailrace	0.8(39)	0.8(44)	0.1(47)	0.5(37)	0.3(60)	0.3(80)	0.5(72)
Tailrace BRZ	14.7(16)	17.9(15)	9.2(17)	13.3(9)	2.4(14)		
Lower Monumental Res.							
Tailrace		1.5(40)			0.3(39)	0.1(38)	0.2(24)
Tailrace BRZ		16.3(16)			1.2(5)		1.0(8)
Little Goose Reservoir							
Tailrace		1.6(40)			0.4(3 1)	0.1(32)	0.3(33)
Tailrace BRZ		28.3(17)			6.4(8)	10.3(8)	1.0(4)

Appendix Table B-2. Indices of northern squawfish abundance from 1990-96 for sampling areas in the lower Columbia and Snake rivers. RKm = river kilometer, BRZ = boat restricted zone.

			Abundance	e Index			
Location, area	1990	1991	1992	1993	1994	1995	1996
Below							
Bonneville Dam							
RKm 71-121			26.8		15.4	14.5	12.2
RKm 122-177			19.7		26.2	17.4	18.7
RKm 178-224			17.9		22.1	14.2	16.4
Tailrace	4.5	5.4	2.7	7.6	2.3	1.8	2.2
Tailrace BRZ	3.0	4.1	2.8	3.1	4.1	1.0	1.3
Bonneville Reservoir							
Forebay	5.5			2.1	2.3	2.3	1.3
Mid-reservoir	15.2			8.5	5.0	7.4	4.9
Tailrace	0.4			0.8	0.5	0.8	0.7
Tailrace BRZ	0.9			0.2	1.1		
The Dalles Reservoir							
Forebay	1.4			1.6	0.7	0.5	0.6
Tailrace	2.7			0.7	0.6	1.5	3.6
Tailrace BRZ	4.4			2.2	1.1	0.7	0.7
John Day Reservoir							
Forebay	1.4	1.3	2.5	1.2	1.4	0.5	0.6
Mid-reservoir	5.2	4.7	6.6	3.2	2.3	1.0	1.1
Tailrace	1.4	1.4	0.2	0.9	0.5	0.6	1.0
Tailrace BRZ	1.6	1.9	1.0	1.4	0.3		
Lower Monumental Res.							
Tailrace		1.3			0.3	0.1	0.1
Tailrace BRZ		0.8			0.1	0.2	0.1
Little Goose Reservoir							
Tailrace		0.7			0.2	< 0.1	0.1
Tailrace BRZ		1.7			0.4	0.6	0.1
Lower Granite Reservoir							
Upper reservoir		1.6			0.5	0.2	0.2

Appendix Table B-3. Indices of northern squawfish consumption of juvenile salmonids from 1990-96 during spring in the lower Columbia and Snake rivers. RKm = river kilometer, BRZ = boat restricted zone.

	Consu	imption Inc	lex (number	of digesti	ive tracts e	examined)	
Location, area	1990	1991	1992	1993	1994	1995	1996
Below							
Bonneville Dam							
RKm 71-121			0.5(102)		0.5(14)	0.5(25)	0.4(23)
RKm 122-177			1.0(189)		1.1(34)	0.2(28)	0.1(43)
RKm 178-224			1.1(126)		1.5(42)	0.7(25)	0.4(33)
Tailrace	1.2(61)		0.5(22)	0.8(75)	3.2(57)	0.8(25)	0.4(29)
Tailrace BRZ	2.7(86)		1.0(77)	1.1(63)	0.6(95)	1.7(17)	0.6(8)
Bonneville Reservoi	r						
Forebay	0.6(153)			0.7(20)	0.2(116)	0.3(88)	0.0(59)
Mid-reservoir	0.0(39)			0.0(14)	0.2(34)	0.0(26)	0.1(17)
Tailrace	0.3(7)			O.O(18)	0.0(19)	0.2(22)	0.0(35)
Tailrace BRZ	2.3(41)						
The Dalles Reservoi	r						
Forebay	0.8(38)			0.1(19)	0.1(22)	0.0(22)	0.0(15)
Tailrace	0.7(27)			0.0(8)	` ´	` ´	` ´
Tailrace BRZ	0.9(50)			0.0(1)			
John Day Reservoir							
Forebay	1.5(38)	1.9(23)	1.9(38)	1.5(1	1) 1.0(11)	1.7(7)	0.7(4)
Mid-reservoir	0.0(6)	0.5(6)	0.0(8)	0.0(2)	0.0(3)	0.0(1)	0.0(3)
Tailrace	1.5(17)	0.9(23)	0.0(9)	2.0(24)	0.3(13)	0.8(13)	0.5(14)
Tailrace BRZ	2.5(60)	1.5(55)	, ,	` ´	0.7(8)		
Lower Monumental	Res.						
Tailrace		0.6(58)			0.7(9)	0.0(2)	0.0(2)
Tailrace BRZ		0.7(12				1.3(7)	0.0(1)
Little Goose Reserv	oir						
Tailrace		0.7(6	8)		1.9(12)	1.4(2)	0.7(7)
Tailrace BRZ		1.2(12	,		, ,	1.6(63) -	-
Lower Granite Rese	rvoir						
Upper-reservoir		0.3(12)	7)		0.6(4	1) 1.2(1	6) 0.2(23)

Appendix Table B-4. Indices of northern squawfish consumption of juvenile salmonids from 1990-96 during summer in the lower Columbia River. RKm = river kilometer, BRZ = boat restricted zone.

	Consur	nption Inc	lex (number	of digestiv	ve tracts ex	xamined)	
Location, area	1990	1991	1992	1993	1994	1995	1996
Below							
Bonneville Dam							
RKm 71-121			0.3(117)		1.8(22	2) 1.5(14)	0.0(9)
RKm 122-177			1.3(136)	-	1.5(32)	0.4(22)	0.0(9)
RKm 178-224			1.9(59)		0.4(32	2) 1.2(20)	0.0(20)
Tailrace	0.5(45)		2.1(43)	1.2(81)	*	9(11)	
Tailrace BRZ	5.5(109)		7.8(147)	1.0(131)	2.1(56)	1.3(16)	3.1(33)
Bonneville Reserve	oir						
Forebay	1.8(139)			0.5(95)	0.3(111)	0.0(96)	0.3(44)
Mid-reservoir	0.0(42)			0.0(3 1			0(19) 0.0(21)
Tailrace	0.0(4)			0.0(14)	O.O(15)	0.8(67)	
Tailrace BRZ	0.8(61)			1.0(23)	3.2(54)		
The Dalles Reserve	oir						
Forebay	1.0(61)			0.0(28)	0.0(27)	0.0(15)	0.0(8)
Tailrace	0.0(46)			0.0(9)	0.8(27)	0.0(41)	0.7(92)
Tailrace BRZ	6.4(50)			0.5(117)	` '	2.2(28)	5.4(3)
John Day Reservoi	r						
Forebay	2.4(16)	3.1(17	0.7(27)	0.6(40)	1.2(57)	2.0(13)	0.4(13)
Mid-reservoir	0.9(7)	0.0(3)	, ,	0.6(10)	• •	0.0(4)	0.0(0)
Tailrace	2.6(25)	0.0(19	,	. , .	, ,	0.6(13)	0.3(19)
Tailrace BRZ	11.7(50)	•	ý 4.6(6 7) 0	•		-	

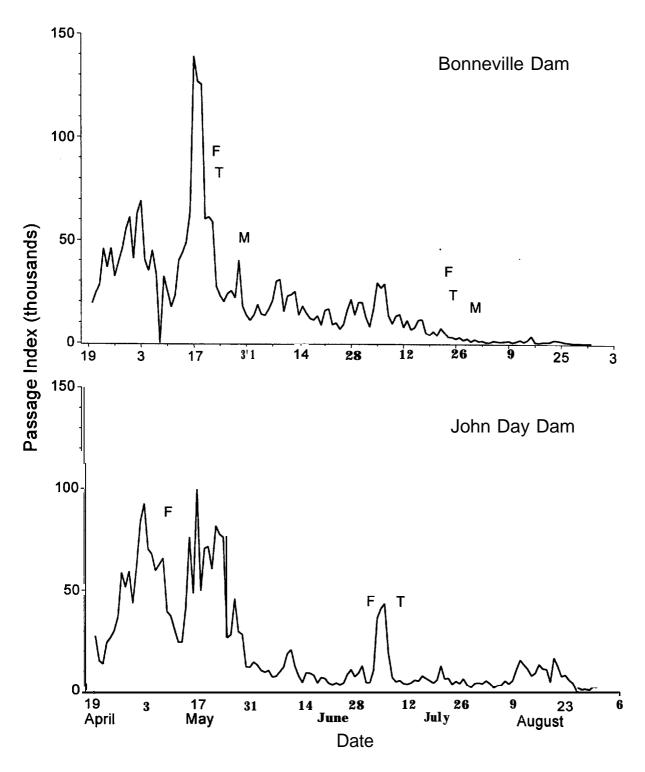
Appendix Table B-5. Indices of northern squawfish predation of juvenile salmonids from 1990-96 during spring in the lower Columbia and Snake rivers. **RKm** = river kilometer, BRZ = boat restricted zone.

			Pre	dation Inc	lex		
Location, area	1990	1991	1992	1993	1994	1995	1996
Below							
Bonneville Dam							
RKm 71-121			10.1		8.0	7.3	4.9
RKm 122-177		40.50	20.5	***	25.5	3.5	1.9
RKm 178-224			33.4		33.2	9.9	6.6
Tailrace	5.5		1.4	6.1	7.4	1.4	0.9
Tailrace BRZ	8.0		2.8	3.5	2.5	1.7	0.8
Bonneville Reservoir							
Forebay	3.3			1.5	0.3	0.7	0.0
Mid-reservoir	0.0			0.0	1.0	0.0	0.5
Tailrace	0.1			0.0	0.0	0.2	0.0
Tailrace BRZ	2.0					1.5	
The Dalles Reservoir							
Forebay	1.1			0.2	0.1	0.0	0.0
Tailrace	1.9			0.0			
Tailrace BRZ	3.9			0.0			
John Day Reservoir							
Forebay	2.1	2.4	4.7	1.9	1.3	0.9	0.4
Mid-reservoir	0.0	2.4	0.0	0.0	0.0	0.0	0.0
Tailrace	1.9	1.3	1.9	1.7	0.2	0.5	0.3
Tailrace BRZ	3.9	2.9	0.9		0.2		
Lower Monumental Res.							
Tailrace		0.8			0.2	0.0	0.0
Tailrace BRZ		0.6				0.3	0.0
Little Goose Reservoir							
Tailrace		0.5			0.4	< 0.1	0.1
Tailrace BRZ		2.0			0.6	1.0	
Lower Granite Reservoir							
Upper-reservoir		0.5			0.3	0.2	0.1

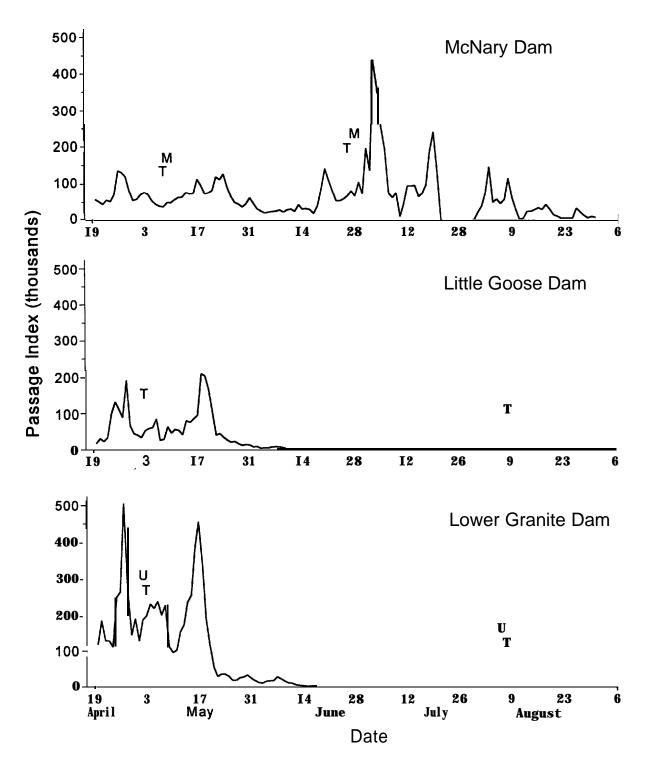
Appendix Table B-6. Indices of northern squawfish predation of juvenile salmonids from 1990-96 during summer in the lower Columbia River. RKm = river kilometer, BRZ = boat restricted zone.

	Predation Index									
Location, area	1990	1991	1992	1993	1994	1995	1996			
Below										
Bonneville Dam										
RKm 71-121			6.0		27.3	14.5	0.0			
RKm 122-177			26.6		34.8	7.0	0.0			
RKm 178-224			37.8		9.5	17.0	0.0			
Tailrace	2.3		5.7	9.1	1.0	1.6	1.3			
Tailrace BRZ	16.4		21.8	3.2	1.3	1.2	4.0			
Bonneville Reservoir										
Forebay	9.9			1.1	0.6	0.0	0.4			
Mid-reservoir	0.0			0.0	0.0	0.0	0.0			
Tailrace	0.0			0.0	0.0	0.6	0.0			
Tailrace BRZ	0.7			0.2	3.5					
The Dalles Reservoir										
Forebay	1.4			0.0	0.0	0.0	0.0			
Tailrace	0.0			0.0	0.5	0.0	2.5			
Tailrace BRZ	27.8			1.1	1.4	1.5	3.8			
John Day Reservoir										
Forebay	3.4	4.0	1.7	0.8	1.6	1.0	0.2			
Mid-reservoir	4.7		0.0	2.0	1.4	0.0	0.0			
Tailrace	3.8		0.0	0.0	0.0	0.4	0.2			
Tailrace BRZ	18.6	5.4	4.6	0.9	0.5					

APPENDIX C	
Timing of 1996 Consumption Index Sampling in Relation to Juvenile Salmonid at Lower Columbia and Snake River Dams	Passage Indices



Appendix Table C-l. Timing of consumption index sampling with respect to juvenile salmonid passage indices at Bonneville and John Day dams in 1996. Sample times for forebay (F), tailrace (T), and areas downstream from Bonneville Dam tailrace (M) are shown.



Appendix Table C-2. Timing of consumption index sampling with respect to juvenile salmonid passage indices at McNary, Little Goose, and Lower Granite dams in 1996. Sample times for tailraces (T), mid-reservoir areas of McNary Reservoir (M), and upper Lower Granite Reservoir (U) are shown.

APPENDIX D

Digestive Tract Contents and Comparison of Fish Diets of Northern Squawfish, Smallmouth Bass, and Walleye in 1996 We examined digestive tract contents of 665 northern squawfish, 870 smallmouth bass, and 112 walleye collected during standardized electrofishing in **1996** (**Appendix Table D-I**). The systemwide frequency of occurrence (%FO) of salmonids in digestive tracts was 6.9% in northern squawfish and 0.8% in smallmouth bass, or approximately 50% lower than in 1994-95 (Zimmerman et al. in press). The %FO of salmonids in walleye was 5.2%. The %FO of salmonids in northern squawfish was highest in Bonneville Dam tailrace during both spring (18.9%) and summer (19.7%). Frequency of occurrence of salmonids in smallmouth bass from all Snake River locations during spring was 3.2%. During summer, the %FO of salmonids in smallmouth bass was greatest below Bonneville Dam tailrace (2.7%) and in John Day Reservoir (1.7%). Systemwide %FO of salmonids in walleye was higher during spring (8.0%) than summer (1.6%).

The composition of prey fishes (by family) found in northern squawfish and smallmouth bass in 1996 was generally similar to 1994-95 (Zimmerman et al. in press). Salmonids comprised the majority (43-83%) of prey fish consumed by northern squawfish, whereas sculpins (Cottus spp.) comprised the majority (62-86%) consumed by smallmouth bass (Appendix Table D-2). The proportion of salmonids found in both northern squawfish and smallmouth bass was greatest in the Snake River. Smallmouth bass utilized prey other than salmonids and sculpins to a far greater extent than northern squawfish, with cyprinids making up the largest proportion of other prey taxa in all three reaches. The proportion of salmonids declined and the proportion of sculpins increased in the diets of northern squawfish and smallmouth bass in 1996. The most common prey fishes utilized by walleye were cyprinids and sandrollers *Percopsis transmontana*. Salmonids comprised a greater proportion of the diets of walleye than of smallmouth bass. No salmonids were found in walleye stomachs downstream from Bonneville Dam; this is probably attributable to a small sample size. Total number of prey fish per predator (all areas combined) was highest (0.59) for walleye, followed by 0.23 prey fish per smallmouth bass and 0.18 prey fish per northern squawfish. Number of prey consumed per predator in 1996 was 33-46% lower than 1994-95 for smallmouth bass and northern squawfish.

Appendix Table D-l. Number of northern squawfish, smallmouth bass, and walleye digestive tracts examined (N) from the lower Columbia and Snake rivers in 1996 that contained food, fish, and juvenile salmonids (Sal).

D : 1			No	Smallmouth bass					Walleye					
Period: or area	N	Food	Fish	Sal N	Food	Fish	Sal	N I	Food	Fish	Sal		j	Re servoir
Spring:														
Below	Bonr	neville												
Dam			99	43	18	8	39	19	6	0	2	2	2	0
Bonnev		Dam												
tailrace		37	15	10	7	6	5	4	0	8	5	4	0	
Bonneville		111	41	10	1	174	120	36	0	10	1	1	0	
The Dalles		15	7	0	0	100	81	27	0	4	3	3	2	
John Day		21	9	2	3	111	84	16	0	26	10	8	2	
Lower Mon.		3	2	2	0	6	5	4	0	0				
Little Goose		7	6	4	4	35	30	13	1	0				
Lower	Gran	ite	23	16	4	3	83	54	10	3	0			
Summer:														
Below	Bonn	neville												
Dam	tailra	ice	38	7	2	0	13	8	6	1	1	1	1	0
Bonnev	ille l	Dam												
tailrace		71	26	19	14	7	4	3	0	5	3	2	0	
Bonneville		88	27	3	0	56	35	12	0	5	4	4	0	
The Dalles		103	20	7	4	55	40	11	0	20	8	7	0	
John D	ay		32	15	2	2	115	89	21	2	31	21	21	1
Lower	Mon		9	2	1	0	24	18	4	0	0			
Little C	Joose	•	6	1	0	0	20	13	2	0	0			
Lower	Gran	ite	3	2	0	0	26	18	7	0	0			

Appendix Table D-2. Sample size and percentage of salmonids, cottids, and other fish families in northern squawfish, smallmouth bass, and walleye digestive tracts containing identifiable fish in three reaches of the lower Columbia and Snake rivers in 1996. Reaches are downstream from Bonneville Dam (DBD), lower Columbia River reservoirs (COL), and lower Snake River reservoirs (SNK).

Sample size,	Northe	ern squa	wfish	Small	mouth	bass	Walleye family		
DBD	COL	SNK	DBD	COL	SNK	DBD	COL	SNK	111.V
Digestive tracts	246	370	50	65	611	194	16	96	0
Total fish	64	37	16	21	134	41	9	57	
Identifiable fish	57	30	12	16	85	35	5	38	
% Salmonidae	63.2	43.3	83.3	6.2	3.5	11.4	0	13.2	
% Cottidae	26.3	56.7	16.7	62.5	85.9	48.6	20.0	42.1	
% Other taxa	10.5	0	0	31.3	10.6	40.0	80.0	44.7	
% Cyprinidae	1.8			12.5	3.5	20.0	40.0	23.7	
% Catostomidae	3.5			6.2	0	2.9	0	2.6	
% Ictaluridae	0			0	1.2	17.1	0	0	
% Percopsidae	5.3			6.2	3.5	0	40.0	3.5	
% Gasterosteidae	0			0	0	0	0	0	
% Centrarchidae	0			0	1.2	0	0	0	
% Percidae	0			6.2	1.2	0	0	0	

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